



Reflection/Review

# Microalgal Technologies: Treatment Strategy for Contaminants of Emerging Concern

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<https://doi.org/10.55295/PSL.12.2025.II1>**Publisher's Note:** UL ZF stays neutral with regard to jurisdictional claims in published maps and institutional affiliations.**Copyright:** © 2025 by the authors. Submitted for possible open access publication under the terms and conditions of the Creative Commons Attribution (CC BY) license (<https://creativecommons.org/licenses/by/4.0/>).**Abstract:**

The increasing presence of contaminants of emerging concern (CEC) in aquatic environments poses significant risks to ecosystems and human health. Traditional wastewater treatment plants (WWTPs) are not equipped to efficiently remove these pollutants, necessitating the development of advanced treatment methods. This article explores the potential of microalgal technologies as a sustainable and cost-effective strategy for CEC removal. Microalgae, such as *Chlorella* sp., *Chlamydomonas* sp., and *Scenedesmus* sp., demonstrate resilience to high concentrations of CEC and offer multiple mechanisms for pollutant removal, including biosorption, bioaccumulation, biodegradation, photodegradation, and volatilization. The article discusses various cultivation methods, including planktonic, immobilized, and mixed biofilm cultures, and highlights the advantages and challenges associated with each. Further research is needed to optimize microalgal cultivation conditions and integration into wastewater treatment processes in order to achieve high CEC removal efficiency. By addressing these challenges, microalgal technologies can enhance the efficiency and sustainability of wastewater treatment systems, contributing to improved water quality and environmental protection.

**Keywords:** Microalgae; wastewater treatment plants; quaternary treatment; *Chlorella* sp.; cultivation methods



## 1. Introduction

Traditionally, water quality management has concentrated on addressing key pollutants such as nutrients, organic matter, suspended solids, heavy metals, and human pathogens (Pal et al., 2014); however, in recent years, growing attention has been directed toward contaminants of emerging concern (CEC). These contaminants pose significant risks to aquatic ecosystems and human health, sparking widespread concern and prompting further research (Tran et al., 2018).

CEC encompass both naturally occurring and artificially synthesized chemicals that are associated with known or uncertain adverse effects on human health and the environment (Ahmed et al., 2021). These substances are characterized by their relative polarity, resistance to natural biodegradation, and potential for bioaccumulation (Kumar et al., 2022). CEC include a wide array of compounds, such as pharmaceuticals and personal care products, endocrine-disrupting compounds, perfluorinated compounds, surfactants, gasoline additives, disinfection by-products, algal and cyanobacterial toxins, organometallic compounds, brominated and organophosphate flame retardants, plasticizers, and nanoparticles.

Most CEC enter waterways through human-mediated routes, including the direct discharge of raw or treated municipal wastewater, landfill leachate, urban and rural runoff, groundwater infiltration, and industrial waste discharge (Tran et al., 2018). Among these, municipal wastewater treatment plants (WWTPs) are the most significant contributors to the continuous discharge of CEC into aquatic environments. The number of specific CEC detected in wastewater discharge ranges from 25 to 200 compounds, a figure likely to increase with advancements in analytical techniques and the industrial use of newly synthesized compounds (Norvill et al., 2016; Tran et al., 2018; Villar-Navarro et al., 2018; Wang et al., 2016).

Conventional WWTPs were originally designed to remove nutrients and organic matter and are not equipped to efficiently eliminate CEC (Lopez et al., 2022). These traditional methods, including processes like waste stabilization ponds, up-flow anaerobic sludge blankets, activated sludge anoxic-aerobic process and sequencing batch reactors, achieve CEC removal rates of only 40–68% (Dubey et al., 2023).

To address these limitations, several advanced technologies have been introduced to enhance the removal of CEC in WWTP effluents, such as chemical precipitation, advanced oxidation process, reverse osmosis, ion exchange, adsorption-based technologies and electroflocculation. However, these methods often incur high costs due to the need for specialized materials like photocatalysts and synthesized adsorbents, as well as fluctuating factors such as material availability, labor, electricity, and recycling costs (Patel et al., 2019). Given these challenges, the development of cost-effective and sustainable treatment methods for CEC removal is imperative. Integrating microalgae into wastewater treatment systems presents a promising, eco-friendly approach (Kumar et al., 2022), with high potential of resources recovery and reuse (e.g. algal biomass) The goal of this paper is to provide a critical review of algal-based technologies for removal of CEC in order to assess their potential and indicate the direction of further research to address their challenges.

## 2. Advantages of microalgal - based bioremediation

The algal-based technologies have emerged as a promising approach for wastewater treatment, offering numerous advantages. Research has shown that various microalgae species, such as *Chlorella* sp., *Chlamydomonas* sp., *Scenedesmus* sp., and *Picocystis* sp., exhibit resistance to a wide range of CEC even at high concentrations. This resilience makes them valuable candidates for removing pollutants from wastewater with elevated contamination levels (Hom-Diaz et al., 2015; Kumar et al., 2022; Prosenc et al., 2021; Quan et al., 2023). Microalgae can efficiently assimilate nutrients present in wastewater, performing bioremediation while significantly reducing the expenses associated with chemical inputs (Kumar et al., 2022). Additionally, this process is environmentally friendly, enhancing CO<sub>2</sub> sequestration and increasing dissolved oxygen (DO) levels in water through the photosynthetic activity of microalgal cells (Qi et al., 2021). Furthermore, microalgae inhibit the growth of harmful bacteria and bloom-forming organisms, thereby reducing microbial loads in aquatic systems (Zhang et al., 2022).



Beyond improving wastewater treatment efficiency, this strategy also supports wastewater valorization by generating valuable biomass. This biomass can be utilized to produce eco-friendly bioproducts, contributing to the circular bioeconomy and promoting sustainable resource management (Kumar et al., 2022).

### 3. Microalgal mechanisms for mitigating CEC

#### 3.1 Biosorption/bioadsorption

Bioadsorption by microalgal cells occurs when contaminants are adsorbed either onto the cell wall components or onto organic substances, such as extracellular polysaccharides (EPS), excreted by the cells into the surrounding environment (Saavedra et al., 2018). This process is non-selective and reversible, allowing various types of CEC to be simultaneously captured on the microalgal cell surfaces (Prosenec et al., 2021). As a passive, non-metabolic interaction, bioadsorption involves the binding of CEC to the negatively charged cell surfaces, which include both the cell wall and secreted substances (collectively termed cell surfaces). These surfaces exhibit a chemical affinity for positively charged contaminants (Xiong et al., 2017).

The ability of CEC to adsorb onto microalgal surfaces depends on its chemical structure. Hydrophobic, cationic contaminants are drawn to the microalgal surface through electrostatic interactions, while hydrophilic CEC tend to be repelled (Xiong et al., 2017). Once a CEC reaches the cell surface, the extent of adsorption is determined by the surface area and chemical composition of the microalgal cells (Norvill et al., 2016). Microalgal cell surfaces contain a variety of functional groups, including carboxyl, hydroxyl and sulfate groups, each with varying affinities for organic and inorganic compounds (Çetinkaya Dönmez et al., 1999).

Bioadsorption involves several chemical processes at the cell surface, including adsorption reactions, ion exchange, surface complexation reactions, chelation, and microprecipitation (Çetinkaya Dönmez et al., 1999). The rate and thermodynamics of the adsorption process are influenced by the physico-chemical properties of the surrounding environment, such as temperature, redox potential and pH.

As bioadsorption is a non-metabolic process, contaminants can bind to both living and non-living microalgal cells, as most cell receptors for CEC remain functional even after the cell has died (Choi & Lee, 2015).

#### 3.2 Bioaccumulation

Bioaccumulation is an active, energy-driven process by which CEC are stored in their original form within microalgal cells. This occurs through the crossing of the cell membrane, where CEC bind to intracellular proteins and other compounds. Unlike bioadsorption, bioaccumulation is slower and occurs exclusively in living cells (Xiong et al., 2018).

Microalgal cells can uptake CEC through three primary mechanisms: passive diffusion, passive-facilitated diffusion and energy-dependent/active uptake across the cell membrane.

In passive diffusion, CEC move across the membrane from areas of high external concentration to areas of low internal concentration without requiring energy expenditure from the cell. This process is facilitated by the hydrophobic nature of the cell membrane, which allows low molecular weight, non-polar, lipid-soluble contaminants to diffuse through. However, polar molecules, high molecular weight compounds, and ions cannot pass through the membrane passively (Xiong et al., 2018).

Passive-facilitated diffusion refers to the process by which CEC cross the cell membrane with the assistance of transporter proteins. These proteins facilitate the influx of polar molecules into the cell (Wilde & Benemann, 1993).

The final mechanism, active transport, involves the movement of CEC across the cell membrane, a process that requires energy expenditure by the cell. In active transport, the compound often moves against a concentration gradient, though this is not always the case. Regardless of the mechanism, bio-uptake is influenced by various factors in the surrounding environment. These include the physico-chemical conditions, such as temperature and



pH, the metabolic state and health of the cell, and the presence of any metabolic inhibitors (Wilde & Benemann, 1993).

### 3.3 Biodegradation (biotransformation)

Biodegradation of CEC by microalgae is one of the most promising pathways for the remediation of these pollutants. Unlike bioadsorption or bioaccumulation, which function primarily as biological filters to concentrate and remove CEC from the surrounding aqueous solution, biodegradation involves the metabolic transformation of complex compounds into simpler, less harmful molecules. This process offers an advantage over bioadsorption and bioaccumulation by addressing the issue of managing CEC-laden microalgal biomass generated during these treatments (Tiwari et al., 2017).

Microalgal biodegradation can occur through two primary mechanisms: metabolic degradation and co-metabolism. In metabolic degradation, the CEC serves as a carbon source or electron donor/acceptor for the microalga, enabling its breakdown, while in co-metabolism, the CEC is degraded by enzymes that catalyze the degradation of other substrates present in the environment (Tiwari et al., 2017).

The process of biodegradation can occur either intracellularly, extracellularly, or through a combination of both. In the intracellular process, the CEC is bio-uptaken by the microalga. Extracellular degradation, on the other hand, occurs when microalgae secrete enzymes into the extracellular polymeric substances (EPS), which act as an external digestive system. The EPS can also serve as a surfactant and emulsifier, increasing the bioavailability of CEC and facilitating their subsequent uptake by the cell. In combination process, initial degradation typically occurs outside the cell, with the breakdown products being further degraded inside the cell (Tiwari et al., 2017; Xiong et al., 2018).

### 3.4 Photodegradation and volatilization

Even if a CEC cannot be bioremediated by microalgae through bioadsorption, bioaccumulation, or biodegradation, microalgae may still contribute to its successful remediation. Two processes—photodegradation and volatilization—can be enhanced by the presence of microalgae or the microalgal treatment system itself (Abo et al., 2016).

Photodegradation of a CEC can occur through two mechanisms: photolysis and photooxidative degradation. Photolysis happens when a contaminant absorbs light, causing direct chemical alteration and subsequent degradation of the compound. In contrast, photooxidative degradation involves the breakdown of the contaminant through interactions with hydroxyl radicals or other oxidants, which are formed from reactions with dissolved organic molecules or nitrate under light exposure (Abo et al., 2016).

Light exposure is essential for photodegradation, but in microalgal treatment systems, light is often absorbed or scattered by the cells, reducing its availability. This can slow down photodegradation. However, the design and operation of these systems can be adjusted to improve light penetration. Microalgae release dissolved organic molecules (DOM) into the water, which can enhance photodegradation. DOM includes compounds like hydrophilic organic acids and humic substances. These molecules can facilitate photodegradation through various mechanisms, such as producing hydroxyl radicals or participating in redox reactions. This process helps break down CEC. (Norvill et al., 2016). For example, studies have shown that microalgal DOM can aid in the removal of pharmaceuticals like ibuprofen through indirect photodegradation (de Wilt et al., 2016).

Volatilization refers to the loss of volatile organic compounds from the liquid phase into the atmosphere. This process depends on the physico-chemical properties of the CEC, such as the Henry's law constant, and the operating conditions of the treatment system, including factors like aeration or agitation rates, temperature, and atmospheric pressure (Tran et al., 2018). In microalgal-based treatment systems, high aeration rates provided by mixing devices (such as paddlewheels, bubble lift columns, and stirrers), along with increased sunlight and higher temperatures compared to conventional wastewater treatment systems, can enhance the removal of volatile CEC (Matamoros et al., 2015).



#### 4. Efficiency of microalgal CEC removal

Several species of microalgae have been tested for their ability to remove CEC, and the choice of species can significantly impact the efficiency of the bioremediation process (Wang et al., 2016).

Microalgae-based systems with optimized conditions, such as high aeration rates, ample sunlight exposure and suitable operating temperatures (whether during warm or cold seasons), can significantly enhance the removal of volatile CEC, including musk fragrances like tonalide and galaxolide (Matamoros et al., 2015).

Among the most frequently studied species for CEC removal are *Chlorella* sp., *Chlamydomonas* sp., and *Scenedesmus* sp., which have been extensively tested in proof-of-concept studies. These species are particularly valued for their robustness and adaptability under stressful environmental conditions. Despite the high diversity of microalgae species, only a few have been sufficiently studied for their potential in bioremediating CEC (Maryjoseph & Ketheesan, 2020). Appendix 1 presents selected CEC and microalgal cultures utilized in various studies, along with their transfer pathways.

Microalgae can be cultivated either individually or combined in both open and closed systems. These cultivation methods are gaining recognition for their potential in secondary and/or tertiary treatment in wastewater treatment plants (Norvill et al., 2016) and for the removal of CEC (Chi et al., 2019; Kurade et al., 2016). The following sections describe three main cultivation methods: planktonic cultures, immobilized cultures and mixed biofilm cultures.

##### 4.1 Planktonic cultures

Open systems are commonly categorized as stabilization ponds and high-rate algal ponds (HRAPs) (Norvill et al., 2016). The main advantages of open systems include their simplicity in construction and operation, along with low operating costs. However, these systems face several challenges, such as susceptibility to contamination by other microorganisms, loss of CO<sub>2</sub> to the atmosphere, and the requirement for large land areas (De Godos et al., 2012). Additionally, their reliance on environmental factors—such as temperature and light—can limit the efficiency of the cultivation process (Matamoros et al., 2016; Matamoros et al., 2015).

Closed photobioreactors (PBRs) offer an alternative to open ponds and come in various configurations, such as tubular, bubble column, air-lift, and flat panel (Torgal, 2016). Unlike open systems, PBRs reduce microbial contamination and allow for the control of variables like pH, temperature, light, and CO<sub>2</sub> concentration (Huang et al., 2017). However, in both open and closed systems, nutrient composition and biological contamination can impact microalgal biomass production and the removal of contaminants of emerging concern (CEC) (Procházková et al., 2014).

##### 4.2 Immobilized cultures

Cell immobilization refers to the restriction of cell mobility through entrapment or attachment to an organic or inorganic water-insoluble solid support (Bouabidi et al., 2019). Common immobilization methods include adsorption onto surfaces, aggregation, encapsulation in semi-permeable membranes, covalent bonding, and entrapment within porous or fibrous polymers. The inherent ability of microorganisms to attach to surfaces often leads to the formation of biofilms, contributing to the immobilization process (Gonçalves et al., 2017).

Immobilized systems offer several advantages, such as maintaining high microalgal concentrations and activity, improving the efficiency of CEC removal, simplifying harvesting, and enhancing resilience to stress, including exposure to toxic contaminants (Zhang et al., 2016). Immobilized microalgae have been successfully used to remove nitrogen, phosphate, and heavy metals from wastewater (Shen et al., 2009). More recently, they have also been employed to remove CEC from wastewater.

##### 4.3 Mixed cultures

Consortia of microorganisms can exhibit greater robustness and stability in response to sudden environmental conditions, as well as protection against invasion by predators and competitor species, compared to individual species (Subashchandrabose et al., 2011). The



presence of bacteria can benefit microalgae growth by removing oxygen and supplying carbon dioxide, altering the surrounding environment and providing essential nutrients and growth factors, such as chelators and phytohormones (Wang et al., 2016).

Several studies have demonstrated enhanced microalgal growth when associated with bacteria, with mutual benefits arising from the excretion of bacterial growth-promoting factors, including microalgal extracellular polymeric substances (EPS) (Cho et al., 2015; Hernandez et al., 2009).

### 5. Challenges and prospects of microalgal-based removal of CEC

Recent advances in microalgal-based bioremediation of CEC present promising, sustainable alternatives to traditional physicochemical processes in wastewater treatment. However, several significant challenges must be addressed before microalgae can be fully implemented as an eco-friendly and effective bioremediation platform for CEC in wastewater. One of the key obstacles is the relatively low performance of microalgae, as different species exhibit varying levels of adaptability and tolerance to CEC toxicity. Future research should focus on isolating and selecting microalgae strains with high growth rates and improved contaminant removal efficiency, as well as enhancing strain performance through methods such as random mutagenesis, genetic engineering, and adaptive laboratory evolution (Arora & Philippidis, 2021).

Another challenge lies in the potential use of microalgal biomass after CEC treatment for food, feed, or biofertilizer applications. Bioaccumulation of CEC in microalgae may hinder such uses, as contaminants could re-enter the food chain and affect the environment (Hena et al., 2020). Therefore, further studies are needed to ensure the safe and sustainable use of microalgal biomass following CEC treatment. Recent research has shown that laccase enzymes can effectively degrade contaminants like Bisphenol A from growth media (Clark et al., 2022). Toxicity in wastewater can also hinder the growth of certain microalgal species, particularly in large-scale real-world applications. The toxicity of wastewater depends on its source and composition (Pittman et al., 2011). Acclimation or adaptation of microalgae to these conditions is a key area of study. Genetic adaptation has been shown to enable microalgae to tolerate high doses of contaminants such as antibiotics, herbicides, and mine waste (García-Balboa et al., 2013). Additionally, microalgae have demonstrated the ability to acclimate to sub-lethal stresses like heavy metals, singlet oxygen, salinity, and high light conditions, often leading to the production of toxic-degrading enzymes (Osundeko et al., 2014; Singh et al., 2020).

A major bottleneck in the application of microalgae for CEC treatment is the lack of suitable cultivation facilities. Conventional systems like open ponds, raceway systems, and high-rate algal ponds are cost-effective and scalable but face limitations, including inefficient light utilization, contamination from predators and heterotrophs, and the need for large land areas (Cheng et al., 2021). To overcome these limitations, closed photobioreactors such as vertical column, tubular, flat plate, membrane, and biofilm-based systems have been developed. However, these systems are more costly to install and maintain, making them less suitable for large-scale, cost-effective wastewater treatment (Sathinathan et al., 2023).

Therefore, an engineering breakthrough is needed to develop economical, efficient, and practical *in situ* treatment systems for microalgal-based CEC wastewater treatment. Furthermore, most studies on microalgal-based CEC bioremediation focus on synthetic water or culture media under laboratory conditions, with only a few studies addressing natural wastewater or surface water. To ensure the real-world applicability of this technology, further research is required to explore the practical use of microalgae for CEC removal in industrial-scale wastewater treatment systems (Kumar & Shukla, 2023).

Also, further research is essential to determine the most effective stage of wastewater treatment for integrating microalgal cultures. Specifically, studies should focus on identifying whether microalgal cultures are most efficient in the secondary, tertiary, or quaternary stages of treatment. Additionally, research should explore the optimal forms and configurations of microalgal cultures, such as planktonic, immobilized, or mixed biofilm cultures, to maximize the removal of CEC. Understanding these factors will help optimize the design and operation of wastewater treatment systems, enhancing their efficiency and sustainability.



## 6. Conclusions

In conclusion, microalgal technologies present a promising and sustainable approach for the treatment of CEC in wastewater. The integration of microalgae into wastewater treatment systems offers numerous advantages, including bio- or photo- degradation of CEC as opposed to just displacement from water, cost-effectiveness, environmental benefits, and the potential for wastewater valorization. Microalgae can effectively remove a wide range of CEC through mechanisms such as biosorption, bioaccumulation, biodegradation, photodegradation, and volatilization.

However, several challenges must be addressed to fully realize the potential of microalgal-based bioremediation. These include optimizing the selection and enhancement of microalgal strains, efficient integration into existing wastewater treatment systems, ensuring the safe use of microalgal biomass post-treatment, and developing economical and efficient cultivation systems. Further research is essential to determine the most effective stage of wastewater treatment for integrating microalgal cultures and to explore the optimal forms and configurations of these cultures.

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**Appendix A** Table A1. Selected CEC and microalgal cultures with removal efficiency and their transfer pathways.

## Appendix 1

**Table 1A.** Selected CECs and microalgal cultures with removal efficiency and their transfer pathways

Contaminants	Microalgal species	Removal efficiency [%]	Mechanisms (pathways)	Reference
17 $\alpha$ -Boldenone	Mixed consortia	82	Not determined	(Zhou et al., 2014)
17 $\alpha$ -Estradiol	Scenedesmus dimorphus	85	Biodegradation	(Zhang et al., 2014)
17 $\alpha$ -Ethinylestradiol	Chlamydomonas reinhardtii	100	Bioadsorption, bio-uptake, biodegradation	(Hom-Diaz et al., 2015)
	Selenastrum capricornutum	60-95		(Bai & Acharya, 2019)
	Nannochloris sp.	60		(Maes et al., 2014)
	Desmodesmus subspicatus	68		
17 $\beta$ -Boldenone	Mixed consortia	75-86	Not determined	(Zhou et al., 2014)
17 $\beta$ -Estradiol	Chlamydomonas reinhardtii	100	Bioaccumulation,	(Hom-Diaz et al., 2015)



	Desmodesmus subspicatus	95	bioadsorption, biodegradation	(Bai & Acharya, 2019) (Huang et al., 2016)
	Selenastrum capricornutum	88		
	Nannochloris sp.	60		
	Chlorella	>92		
7-amino cephalosporanic acid	Chlorella sp.	100	Photodegradation, bioadsorption	(Guo et al., 2016)
	Chlamydomonas sp.	100		
	Mychonastes sp.	100		
Acetaminophen	Mixed consortia	99	Not determined	(Matamoros et al., 2016)
Alprazolam	Mixed consortia	87	Not determined	(Hom-Diaz et al., 2017a, 2017b)
Amitriptyline	Chlorella sorokiniana	68	Bioadsorption	(Gojkovic et al., 2019)
	Chlorella vulgaris	42		
	Chlorella saccharophila	92		
	Coelastrum sp.	60		
	Coelastrum astroideum	68		
	Desmodesmus sp.	37-47		
	Scenedesmus sp.	85		
	Scenedesmus obliquus	69		
Amoxicillin	Chlorella pyrenoidosa	77	Not determined	(Li et al., 2015)
	Microcystis aeruginosa	18-31		(Liu et al., 2015)
Atenolol	Mixed consortia	85-98	Not determined	(Hom-Diaz et al., 2017a, 2017b)
Azithromycin	Mixed consortia	89	Not determined	(Hom-Diaz et al., 2017a, 2017b)



Biperiden	Chlorella sorokiniana	35	Bioadsorption	(Gojkovic et al., 2019)
	Chlorella vulgaris	93		
	Chlorella saccharophila	89		
	Coelastrum astroideum	9		
	Desmodesmus sp.	41-71		
	Scenedesmus sp.	53		
	Scenedesmus obliquus	48		
Bisphenol A	Chlorella pyrenoidosa	43	Bio-uptake, biodegradation	(Fu et al., 2023)
	Desmodesmus sp. WR1	26		(Wang et al., 2017)
	Chlamydomonas mexicana	24		(Ji et al., 2014)
	Chlorella vulgaris	24		
Bupropion	Chlorella sorokiniana	60	Bioadsorption	(Gojkovic et al., 2019)
	Chlorella vulgaris	82		
	Chlorella saccharophila	88		
	Coelastrella sp.	89		
	Coelastrum astroideum	94		
	Desmodesmus sp.	86-90		
	Scenedesmus sp.	70		
	Scenedesmus obliquus	95		
Caffeine	Mixed consortia	99	Biodegradation	(Matamoros et al., 2016)
	Mixed consortia	26-81		(Gojkovic et al., 2019)
Carbamazepine	Mixed consortia	4-15	Bioadsorption, biodegradation	(Zhou et al., 2014)
	Mixed consortia	20		(Matamoros et al., 2016)



	Chlamydomonas mexicana	35		(Xiong et al., 2017b)
	Scenedesmus obliquus	35		
	Chlorella sorokiniana	10-30		(de Wilt et al., 2016)
	Nannochloris sp.	20		(Bai & Acharya, 2016)
	Desmodesmus sp.	71		(Gojkovic et al., 2019)
Carbendazim	Mixed consortia	14-30	Not determined	(Zhou et al., 2014)
Cefradine	Chlorella pyrenoidosa	23	Not determined	(Li et al., 2015)
Cefradine	Chlorella pyrenoidosa	76	Not determined	(Chen et al., 2015)
Ciprofloxacin	Chlorella vulgaris	89,9	Biodegradation, bioadsorption, photodegradation	(Al-Mashhadani and Al-Mashhadani, 2023)
	Chlamydomonas sp.Tai-03	100		(Xie et al., 2020)
	Mixed consortia	74-79		(Zhou et al., 2014)
	Mixed consortia	20-30		(Hom-Diaz et al., 2017a, 2017b)
	Chlamydomonas mexicana	13-56		(Xiong et al., 2017b)
	Nannochloris sp.	100		(Bai & Acharya, 2016)
	Dictyosphaerium sp.	11		(Gentili & Fick, 2017)
Clarithromycin	Mixed consortia	100	Not determined	(Zhou et al., 2014)
Climbazole	Mixed consortia	30-70	Biodegradation	(Zhou et al., 2014)
	Scenedesmus obliquus	88		(Pan et al., 2018)
Clofibric acid	Mixed consortia	0-30	Not determined	(Zhou et al., 2014)



Clomipramine	Chlorella sorokiniana	96	Bioadsorption	(Gojkovic et al., 2019)
	Chlorella vulgaris	100		
	Chlorella saccharophila	100		
	Coelastrella sp.	34		
	Desmodesmus sp.	29-42		
	Scenedesmus sp.	73		
	Scenedesmus obliquus	78		
Codeine	Chlorella sorokiniana	50	Biodegradation, photodegradation	(Gojkovic et al., 2019)
	Chlorella vulgaris	57		
	Chlorella saccharophila	42		
	Coelastrella sp.	46		
	Coelastrum astroideum	72		
	Desmodesmus sp.	37-80		
	Scenedesmus sp.	33		
	Scenedesmus obliquus	59		
Diazinon	Chlorella vulgaris	94	Biodegradation	(Kurade et al., 2016)
Diclofenac	Picocystis sp. CINS 23	>90	Bioadsorption, bioaccumulation, biodegradation	(Ali et al., 2022)
	Mixed consortia	92		(Matamoros et al., 2016)
	Chlorella sorokiniana	40-60		(de Wilt et al., 2016)
	Chlorella sorokiniana	30		(Escapa et al., 2015)
	Chlorella vulgaris	21		(Villar-Navarro et al., 2018)
	Mixed consortia	55		
Diltiazem	Mixed consortia	72-77	Not determined	(Hom-Diaz et al., 2017a, 2017b)



Di-n-butyl phthalate	Chaetoceros muelleri	22,5	Biodegradation	(Chi et al., 2019)
	Cylindrotheca closterium	91,4		
Diphenhydramine	Chlorella sorokiniana	73	Biodegradation	(Gojkovic et al., 2019)
	Chlorella vulgaris	98		
	Chlorella saccharophila	93		
	Coelastrella sp.	87		
	Coelastrum astroideum	87		
	Desmodesmus sp.	88-92		
	Scenedesmus sp.	86		
	Scenedesmus obliquus	85		
Enrofloxacin	Mixed consortia	75-77	Not determined	(Zhou et al., 2014)
Erythromycin	Mixed consortia	63-86	Not determined	(Zhou et al., 2014)
	Mixed consortia	85		(Hom-Diaz et al., 2017a, 2017b)
Estriol	Scenedesmus dimorphus	85	Biodegradation	(Zhang et al., 2014)
Estrone	Mixed consortia	85	Biodegradation	(Zhou et al., 2014)
	Scenedesmus dimorphus	85		
Flecainide	Chlorella sorokiniana	71	Photodegradation	(Gojkovic et al., 2019)
	Chlorella vulgaris	100		
	Chlorella saccharophila	100		
	Coelastrella sp.	52		
	Coelastrum astroideum	66		
	Desmodesmus sp.	72-96		
	Scenedesmus sp.	40		



	Scenedesmus obliquus	93		
Fluconazol	Desmodesmus sp.	33	Bioadsorption	(Gojkovic et al., 2019)
Fluoxastrobin	Synechococcus sp.	Not determined	Bioadsorption	(Stravs et al., 2017)
Fluxonazole	Mixed consortia	25	Not determined	(Zhou et al., 2014)
Hydrochlorothiazide	Mixed consortia	44-84	Not determined	(Hom-Diaz et al., 2017a, 2017b)
Hydroxyzine	Chlorella sorokiniana	76	Biodegradation	(Gojkovic et al., 2019)
	Chlorella vulgaris	93		
	Chlorella saccharophila	93		
	Coelastrella sp.	80		
	Coelastrum astroideum	96		
	Desmodesmus sp.	87-100		
	Scenedesmus sp.	73		
	Scenedesmus obliquus	95		
Ibuprofen	Mixed consortia	99	Bio-uptake, biodegradation	(Matamoros et al., 2016)
	Mixed consortia	98		(Hom-Diaz et al., 2017a, 2017b)
	Chlorella sorokiniana	100		(de Wilt et al., 2016)
	Nannochloris sp.	40		(Bai and Acharya, 2016)
	Navicula sp.	60		(Ding et al., 2017)
Ketoprofen	Mixed consortia	36-85	Not determined	(Hom-Diaz et al., 2017a, 2017b)
Kresoxim-methyl	Mixed consortia	Not determined	Not determined	(Stravs et al., 2017)



Levofloxacin	Scenedesmus obliquus	93,4	Biodegradation, bioaccumulation, bioadsorption	(Xiong et al., 2017a)
	Chlorella vulgaris	10-90		
Lincomycin	Mixed consortia	80	Not determined	(Zhou et al., 2014)
Lorazepam	Mixed consortia	30-60	Not determined	(de Wilt et al., 2016)
Memantine	Chlorella sorokiniana	87	Bioadsorption, biodegradation	(Gojkovic et al., 2019)
	Chlorella vulgaris	100		
	Chlorella saccharophila	100		
	Coelastrella sp.	78		
	Coelastrum astroideum	73		
	Desmodesmus sp.	44-86		
	Scenedesmus sp.	92		
	Scenedesmus obliquus	86		
Metoprolol	Chlorella sorokiniana	100	Biodegradation	(de Wilt et al., 2016)
	Dictyosphaerium sp.	99		(Gentili & Fick, 2017)
	Chlamydomonas reinhardtii	99		(Stravs et al., 2017)
Metronidazole	Chlorella vulgaris	100	bioadsorption	(Hena et al., 2020)
Mitrazapine	Chlorella sorokiniana	63	Bioadsorption, biodegradation	(Gojkovic et al., 2019)
	Chlorella vulgaris	69		
	Chlorella saccharophila	80		
	Coelastrella sp.	70		
	Coelastrum astroideum	67		
	Desmodesmus sp.	55-85		
	Scenedesmus sp.	77		
	Scenedesmus obliquus	62		



Naproxen	Mixed consortia	89	Biodegradation	(Matamoros et al., 2016)
	Mixed consortia	10-70		(Hom-Diaz et al., 2017a, 2017b)
Nonylphenol	Chlorella pyrenoidosa	48	biodegradation	(Feng et al., 2022)
Norfloxacin	Mixed consortia	41-53	Not determined	(Zhou et al., 2014)
Norgestrel	Chlorella pyrenoidosa	60	Biodegradation	(Peng et al., 2014)
	Scenedesmus obliquus	95		
Ofloxacin	Mixed consortia	43-52	Not determined	(Zhou et al., 2014)
	Mixed consortia	66		(Hom-Diaz et al., 2017a, 2017b)
Orphenadrine	Chlorella sorokiniana	82	Bioadsorption	(Gojkovic et al., 2019)
	Chlorella vulgaris	100		
	Chlorella saccharophila	98		
	Coelastrella sp.	78		
	Coelastrum astroideum	66		
	Desmodesmus sp.	75-82		
	Scenedesmus sp.	79		
	Scenedesmus obliquus	95		
Oxytetracycline	Picocystis sp. CINS 23	100	Bioadsorption, bioaccumulation, biodegradation	(Ali et al., 2022)
Paracetamol	Mixed consortia	88-94	Biodegradation, photodegradation	(Zhou et al., 2014)
	Chlorella sorokiniana	100		(de Wilt et al., 2016)
	Chlorella sorokiniana	41-69		(Escapa et al., 2015)



Paroxetine	Mixed consortia	99	Not determined	(Hom-Diaz et al., 2017a, 2017b)
Phthalic acid esters	Chaetoceros muelleri	95,5	Biodegradation	(Chi et al., 2019)
	Cylindrotheca closterium	97,7		
Progesterone	Mixed consortia	83-87	Biodegradation	(Zhou et al., 2014)
	Chlorella pyrenoidosa	95		(Peng et al., 2014)
	Scenedesmus obliquus	95		
Roxithromycin	Mixed consortia	87-94	Not determined	(Zhou et al., 2014)
Salicylic acid	Mixed consortia	97	Bio-uptake, biodegradation	(Zhou et al., 2014)
	Nannochloris sp.	60		(Bai & Acharya, 2019)
	Mixed consortia	33		(Hom-Diaz et al., 2017a, 2017b)
	Chlorella sorokiniana	73		(Escapa et al., 2015)
	Mixed consortia	90		(Villar-Navarro et al., 2018)
	Chlorella sorokiniana	93-98		(Escapa et al., 2015)
	Chlorella vulgaris	25		(Escapa et al., 2017)
	Scenedesmus obliquus	93		
Salinomycin	Mixed consortia	71-79	Not determined	(Zhou et al., 2014)
Sulfadiazine	Mixed consortia	52-75	Not determined	(Zhou et al., 2014)
Sulfadimethoxine	Mixed consortia	56-78	Not determined	(Zhou et al., 2014)
Sulfamethazine	Mixed consortia	18-48	Not determined	(Zhou et al., 2014)
Sulfamethoxazole	Nannochloris sp.	32	Biodesorption, biodegradation,	(Bai & Acharya, 2019)



	Nannochloris sp.	40	photodegradation	(Bai & Acharya, 2016)
	Mixed consortia	0-18		(Stravs et al., 2017; Zhou et al., 2014)
Sulfapyridine	Mixed consortia	98	Not determined	(Zhou et al., 2014)
Testosterone	Mixed consortia	100	Not determined	(Zhou et al., 2014)
Tetrabromobisphenol A	Chlorella sphaericum & Scenedesmus quadricauda	98	Biodegradation	(Peng et al., 2014)
Tetracycline	Chlorella sp.	68	Bioadsorption, photodegradation	(Suárez-Martínez et al., 2022)
	Scenedesmus quadricauda	48,84		(Daneshvar et al., 2018)
	Chlorella vulgaris	69		(De Godos et al., 2012)
Thiamethoxam	Chlorella sp. TXH	97,5	Biodegradation, bioadsorption, bioaccumulation	(Quan et al., 2023)
Thiamphenicol	Chlorella sp. L38	77,7	Biodegradation, bioaccumulation, and biosorption	(Song et al., 2020)
	Chlorella sp. UTEX1602	87,3		
Tramadol	Dictyosphaerium sp.	57	Bio-uptake, biodegradation	(Gentili & Fick, 2017)
	Scenedesmus obliquus	91		(Ali et al., 2018)
	Chlorella vulgaris	51		(Gojkovic et al., 2019)
	Desmodesmus sp.	14-45		
Trichlorfon	Chlamydomonas reinhardtii	100	Biodegradation	(Wan et al., 2020)
Triclocarban	Mixed consortia	81-99	Not determined	(Zhou et al., 2014)
Triclosan	Mixed consortia	31-58		(Zhou et al., 2014)



	Mixed consortia	95	Biodegradation, photodegradation	(Matamoros et al., 2016)
	Nannochloris sp.	100		(Xiong et al., 2017b)
	Nannochloris sp.	72		(Bai & Acharya, 2016)
	Chlorella pyrenoidosa	77		(Wang et al., 2013)
	Microcystis aeruginosa	46		(Wang et al., 2016)
	Nannochloris sp.	100		(Rühmland et al., 2015)
Trihexyphenidyl	Chlorella sorokiniana	40	Bioadsorption	(Gojkovic et al., 2019)
	Chlorella vulgaris	100		
	Chlorella saccharophila	95		
	Coelastrum astroideum	54		
	Desmodesmus sp.	63-73		
	Scenedesmus sp.	49		
	Scenedesmus obliquus	60		
Trimethoprim	Mixed consortia	0-37	Not determined	(Zhou et al., 2014)
	Chlorella sorokiniana	40-60		(de Wilt et al., 2016)
	Dictyosphaerium sp.	<4		(Gentili & Fick, 2017)
	Chlorella sorokiniana	60		(Escapa et al., 2015)
Tylosin	Mixed consortia	75	Not determined	(Zhou et al., 2014)



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